Picocyanobacterial community structure of freshwater lakes and the Baltic Sea revealed by phylogenetic analyses and clade-specific quantitative PCR

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Phylogenetic relationships among picocyanobacteria from the Syn/Pro clade *sensu* Sánchez-Baracaldo *et al.* (2005) were determined using small subunit (ssu) rDNA sequences from novel culture isolates together with environmental samples from the Baltic Sea and seven freshwater lakes. The picocyanobacterial community comprised members of previously identified clades and of two previously undescribed clades. The number of well-supported clades suggests that freshwater picocyanobacterial communities encompass much greater diversity than is found in marine systems. To allow the quantification of community structure and temporal succession, clade-specific ssu rDNA TaqMan assays were designed and implemented. These assays were used to assess picocyanobacterial community structure in two lakes over an annual cycle in 2003/4, and in a small number of Baltic Sea samples collected in July 2003. In the lake-water samples, picocyanobacteria were found to be scarce during most of the year, with members of each clade reaching their peak abundance over a relatively short period during the summer (June to September), although representatives of the *Cyanobium* clade also developed an autumn peak extending towards the end of October. All four freshwater clades were present in the Baltic Sea, but their distribution was patchy over relatively short spatial scales. The use of molecular tools for describing and quantifying community structures reveals previously unexplored complexity in the phytoplankton and will facilitate the development of a more sophisticated understanding of community dynamics at the base of the food chains in lakes.

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INTRODUCTION

Picoplanktonic cyanobacteria contribute between 5 and 80 % of the total primary production in lakes, depending on the season, water chemistry and hydrography (Stockner *et al.*, 2000). For such an important group of organisms our current understanding of their ecology, diversity and taxonomy in freshwaters is still limited. Studies of picocyanobacterial diversity and community structure in lakes are needed to improve our understanding of the dominant ecological processes at the base of the food web. There is convincing evidence for both marine and freshwater environments that individual water bodies can

support a diverse community of picoplanktonic cyanobacteria (e.g. Stockner *et al.*, 2000). Autotrophic picoplankton (APP), in marine environments, is dominated by a single clade of picocyanobacteria encompassing diverse strains assigned to the genera *Synechococcus* and *Prochlorococcus* (Sánchez-Baracaldo *et al.*, 2005), whereas in freshwater environments the APP community is dominated by organisms generally described as *Synechococcus*-type (Weisse, 1993; Padisák *et al.*, 2003).

Picocyanobacteria lack sufficient readily discernible morphological features to provide a robust taxonomic separation into distinct genera and species, and differences based on pigmentation (e.g. Postius & Ernst, 1999) can be misleading (Ernst *et al.*, 2003). In such morphologically depauperate groups, knowledge of phylogenetic relationships is crucial to understanding lineage diversification and is therefore useful when studying community structures. Most phylogenetic studies of picocyanobacterial communities have focused on the relationships among marine picoplankton (i.e. *Synechococcus* and *Prochlorococcus*) using sequence data derived from the small subunit (ssu) rDNA,

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Abbreviations: APP, autotrophic picoplankton; CWP, Cotswold Water Park; ITS-1, rDNA internal transcribed spacer; ssu, small subunit; Syn/Pro, *Synechococcus*/*Prochlorococcus*/*Cyanobium* clade.

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the rDNA internal transcribed spacer (ITS-1), the phycobiliprotein encoding genes (*cpc* and *cpe*) and others (Urbach *et al.*, 1998; Roca *et al.*, 2002; Ting *et al.*, 2002; Fuller *et al.*, 2003; Steglich *et al.*, 2003). A smaller number of studies have looked at the phylogenetic relationships of non-marine picocyanobacteria and have been primarily based on ssu rDNA, rDNA ITS-1 and *cpcBA*-IGS (intergenic spacer) sequences (Ernst *et al.*, 2003; Crosbie *et al.*, 2003). Although ssu rDNA sequences seem to resolve phylogenetic relationships within relatively closely related strains, they do not provide well-supported resolution at deep branches (Honda *et al.*, 1999; Litvaitis, 2002). However, because of the amount of ssu rDNA sequence data already publicly available in GenBank, this locus represents a useful resource for determining the phylogenetic and taxonomic relationships of sequences derived from organisms that have not yet been isolated into culture. A recent phylogenomic study, in which a backbone phylogenetic tree was calculated using 36 concatenated slowly evolving genes from 14 cyanobacterial genomes and a larger number of taxa with a more limited amount of data (ssu rDNA, *rpoC* and morphology), resolved the deep branches within the Cyanobacteria (Sánchez-Baracaldo *et al.*, 2005). This study suggested that the clade containing the picocyanobacteria, i.e. the *Synechococcus*/*Prochlorococcus*/*Cyanobium* (Syn/Pro) clade, branches deep in the Cyanobacteria, just after *Gloeobacter*, the deepest branch. Character evolution studies, using parsimony reconstruction, also suggest that the early Earth cyanobacteria evolved in terrestrial and/or freshwater environments (Sánchez-Baracaldo *et al.*, 2005), as opposed to marine environments as previously postulated (Honda *et al.*, 1999; Crosbie *et al.*, 2003). It is likely that lineages within the Syn/Pro clade exhibit some morphological and ecological traits present in the earliest cyanobacterial lineages, such as small cell diameter and free-living planktonic habit (Sánchez-Baracaldo *et al.*, 2005).

Long-term monitoring in the mesotrophic Lake Constance indicates potential shifts in the composition of spring and summer picocyanobacterial communities (Gaedke & Weisse, 1998); however, the lack of distinctive morphologies for different taxa within the genus *Synechococcus* has allowed only limited differentiation within such communities (Postius & Ernst, 1999). Phylogenetic studies have shown that *Synechococcus* is polyphyletic, i.e. is not a natural taxon (Honda *et al.*, 1999; Wilmotte & Herdman, 2001; Robertson *et al.*, 2001), and the continuing use of this generic designation in ecology can lead to confusion and a lack of consistency between studies. The recent application of molecular ecology techniques has revealed substantial diversity in freshwater picocyanobacterial communities (Postius & Ernst, 1999; Roca *et al.*, 2002; Ernst *et al.*, 2003; Becker *et al.*, 2004). The use of Taq nuclease assays has allowed sensitive detection of genotypes in natural microbial communities, in which single ecotypes, based on ssu rDNA ITS-1 sequences, have been monitored in both space and time in Lake Constance

(Becker *et al.*, 2000, 2002, 2004, 2007). Quantitative real-time PCR has also allowed quantification of microcystin-encoding genes and cell concentrations in temperate lakes (Vaitomaa *et al.*, 2003; Rinta-Kanto *et al.*, 2005; Schober & Kurmayer, 2006), detection of toxic *Nodularia* in the Baltic Sea (Koskenniemi *et al.*, 2007) and differences in distribution of *Prochlorococcus* ecotypes in the Sargasso Sea and Atlantic Ocean (Ahlgren *et al.*, 2005; Johnson *et al.*, 2006). Clade-specific oligonucleotide probes have revealed structuring of the *Synechococcus* community within a stratified water column in the Red Sea (Fuller *et al.*, 2003). Furthermore, phylogenetically informed use of molecular techniques, such as dot-blot hybridization and fluorescence *in situ* hybridization, has revealed the distribution of ecotypes and lineages of marine *Prochlorococcus* and *Synechococcus* in the Mediterranean Sea and Atlantic Ocean (Garczarek *et al.*, 2007; Zwirgmaier *et al.*, 2007).

To better understand freshwater picocyanobacterial diversity, we have used a phylogenetic approach combined with quantitative Taq nuclease assays (TaqMan PCR) to assess community structure in seven English post-glaciation and man-made lakes. Phylogenetic analyses of ssu rDNA sequences from the marine *Synechococcus* and *Prochlorococcus* lineages, novel cultured freshwater isolates and ssu rDNA clone libraries derived from environmental DNA were used to identify well-supported clades of freshwater picocyanobacteria. Clade-specific TaqMan quantitative PCR assays were used to assess the abundance of the members of four picocyanobacterial clades in freshwater APP communities. Samples for these analyses were collected from the Baltic Sea in the summer of 2003 and over an annual cycle from two lakes in the UK. We show that individual water bodies can support a diverse range of picocyanobacteria and that the community structure is temporally and spatially variable. We also show that the brackish Baltic Sea supports a picocyanobacterial community that has elements of both the freshwater and marine floras and that it is patchy in structure over relatively small spatial scales.

METHODS

Cultures. Cultures of Baltic Sea picocyanobacteria were established from samples collected from the upper mixed layer of the water column in June and July 1998 (Table 1). Samples were pre-filtered through 20 µm mesh plankton netting, then cells were concentrated onto 0.6 µm polycarbonate filters (Nucleopore) and either smeared onto BG11 culture medium (Rippka *et al.*, 1979) solidified with 0.5 % (w/v) agarose or suspended in 1 ml of either BG11 or filter-sterilized seawater in 2 ml screw-cap tubes. Cells on solidified medium were stored under black netting on the open bench until return to the UK. Liquid cell suspensions were initially stored in tanks of running surface seawater and subsequently 50 µl aliquots of 10⁰, 10⁻² and 10⁻⁴ dilutions in BG11 were spread onto BG11 agarose. In the laboratory all plates were incubated at 15 °C under an incident irradiance of 50 µmol m⁻² s⁻¹ with a 12:12 light/dark cycle. Picocyanobacteria were streaked to single colonies to ensure clonality and then maintained in liquid culture. Cultures of picocyanobacteria

Table 1. Picocyanobacteria cultures and clones and their sampling locations

Sampling locations: bs/BS, Baltic Sea; pp/PP, Priest Pot; ew/EW, Esthwaite Water; ap/AP, Abbot's Pool; cwp/CWP, Cotswold Water Park lakes. The number of times each sequence was found is shown in parentheses: where there is no number the sequence was encountered only once.

Sampling location	Sampling dates*	Environmental ssu rDNA sequences		ssu rDNA sequences from cultures§
		Short†	Long‡	
Baltic Sea (55°14'–59°04'N, 13°35'–20°06'E)	24/06–02/07/1998	bsA1-5 (× 11), bsA3-2, bsA3-6, bsL4-3 (× 2)		B9801-type (× 1), B9802-type (× 4), B9803-type (× 1)
Baltic Sea (54°28'–59°29'N, 10°43'–15°48'E)	01–11/07/1998	bsA1-5 (× 5), bsA3-2 (× 3); bsL4-3, bsL28-2		9801-type (× 4), 9802-type (× 5), 9803-type (× 2)
Baltic Sea (58°10'–58°35'N, 18°34'–19°60'E)	23–25/06/1999	bsA1-5 (× 3), bsL4-3		
Priest Pot (54°22'N 02°59'W)	18/06/2002	pp9-1 (× 4)		
Esthwaite Water (54°21'N 02°58'W)	22/07/2003		EW7-2	EW2, EW8 and EW15II
Abbots Pool (51°27'N 02°40'W)	22/07/2002		AP10-7; AP10-8	
CWP 9 (51°39'N 01°53'W)	17/06/2002– 18/07/2002	cwp9-1 (× 4); cwp9.2-2	CWP9.2-7; CWP9.2-33	
CWP 31 (51°39'N 01°57'W)	16/04/2002	cwp31		
CWP 123 (51°41'N 01°43'W)	12/08/2002		CWP123.2-1; CWP123.2-2	
CWP 124 (51°41'N 01°43'W)	18/07/2002	cwp124.3	CWP124.2-1; CWP124.2-10	

*Day/month/year.

†Short sequences were 362–424 bp in length and were derived from amplification products generated using primer pair 16S.19F and 16S.402R (Table 2).

‡Long sequences were 1362–1430 bp in length and were derived from amplification products generated using primer pairs Syn3 and Syn5 (Table 2) or Cyano5' and Cyano3' (Beard *et al.*, 1999).

§Sequences from cultures were 1405–362 bp in length and were derived from amplification products generated using primer pairs Cyano5' and Cyano3' (Beard *et al.*, 1999) or 16S.19F and 16S.402R or CYA359F and CYA781(a)/CYA781(b) (Nübel *et al.*, 1997).

II Cultures isolated in June and October 2000.

isolated from Esthwaite Water (English Lake District) were obtained from Dr Jackie Parry, University of Lancaster.

Samples of natural communities. Picocyanobacteria in pooled samples from the upper mixed layer of the Baltic Sea collected in June 1999 were pre-filtered through 20 µm plankton netting, and then concentrated on to 0.45 µm pore-size membrane filters (Gelman Supor 450) until the filters blocked; filters were stored in 1 ml DNA extraction buffer (10 mM Tris, 0.5 mM NaCl, 100 mM EDTA) at –20 °C. For all other environmental samples (for sampling locations see Tables 1 and 4) surface water was prefiltered sequentially through 30 µm mesh plankton netting and 5 µm pore-size cellulose nitrate membranes (Whatman) prior to collecting cells from a known volume of water (380 ml to 5 l) on 0.45 µm pore-size cellulose acetate membranes (Pall Corporation). Cells immobilized on the membrane filters were rinsed *in situ* with 1 ml DNA extraction buffer and stored at –20 °C.

DNA extraction. DNA from picocyanobacterial cultures was extracted using lysozyme and proteinase K lysis followed by phenol/chloroform/isoamyl alcohol extraction as described by Beard *et al.* (1999). DNA from microbial communities stored frozen on filters in DNA extraction buffer was isolated by phenol/chloroform extraction (Kerkhof & Ward, 1993; Voytek & Ward, 1995). For quantitative PCR

(see below) a 4 mm diameter punch was used to excise one or more 50.3 mm² sections from filters. Filter fragments were incubated with occasional shaking in 400 µl of 5% (w/v) Chelex-100 (sodium form; 100–200 mesh: Bio-Rad) for 30 min at 100 °C, vortexed at high speed and the debris pelleted by centrifugation for 2 min (Becker *et al.*, 2002). DNA-containing supernatants were stored frozen in aliquots at either –20 °C or –80 °C.

Characterization of ssu rDNA sequences. Near-complete ssu rDNA sequences (1430 bp) were amplified using the primer pair Cyano 5' and Cyano 3' (Beard *et al.*, 1999). On the basis of initial results, specific freshwater picocyanobacterial primers, Syn5 and Syn3 (Table 2), were designed to reduce amplification of non-target sequences while still generating near full-length products (1362 bp). For amplification of shorter ssu rDNA fragments (362 bp), cyanobacterial specific primers were used, either 16S.19F and 16S.409R (this study; Table 2) or CYA359F and CYA781R(a)/CYA781(b) (Nübel *et al.*, 1997). Amplification reactions (50 µl) contained 10 mM Tris/HCl (pH 9.0), 1.5 mM MgCl₂, 50 mM KCl, 0.2 mM of each dNTP, 0.5 µM of each primer, 1 µl of 75 µg BSA ml^{–1}, 1 unit Supertaq DNA polymerase (HT Biotechnology) and 1 µl genomic DNA. Cycling conditions for long fragments were: initial denaturation at 94 °C for 3 min; 25 cycles of 94 °C for 45 s, 50 °C for 1 min, and 72 °C for 2 min; final extension at 72 °C for 10 min.

Table 2. Oligonucleotide primers and probes used in this study to amplify and quantify *ssu rDNA* targets

Mismatches with non-target freshwater and marine *Synechococcus* sequences are underlined.

	Sequence (5'–3')	<i>T_m</i>	Target organisms
Primers			
Syn3	TACGACTTCACCCAGTCA(CT)CAGCC		Cyanobacteria
Syn5	CAGGATGAACGCTGGCGGC(CT)(GC)TGC		Cyanobacteria
16S.19F	AAG CCT GAC GGA GCA ACG CC	63	<i>Synechococcus</i>
16S.409R	GGT ATC TAA TCC CTT TCG CTC C	60	<i>Synechococcus</i>
CLF	CAGGCGGCCTTGTAAAGTCTGT	60	<i>Cyanobium gracile</i> cluster/clade I
CLR	ACACTCAAGCCTTGTAGTTTCCAT	61	<i>Cyanobium gracile</i> cluster/clade I
CILF	GCAGGCGGCCTCGAA	59	Clade II
CILR	CCCTACCACACTCTAGCCTCGTAG	59	Clade II
CIILF	CAGGCGGTCTTGTAAAGTCTGC	59	Clade III
CIILR	TCTTGTAGTTTCCACCGCGCA	61	Clade III
CIV.F	TCCGCAGGCGGTTTACA	60	Clade IV
CIV.R	CCCCTACCACACTCGAGTCTTAC	58	Clade IV
Probes			
P100	Fam-AAGCGTGGAGCTTAA-NFQ	70	Clades I, II, III
P104	Fam-AAGCGTGGAGCTCAA-NFQ	69	Clade IV

Cycling conditions for shorter fragments were: initial denaturation at 94 °C for 3 min; 30 cycles of 94 °C for 1 min, 60 °C for 1 min and 72 °C for 1.5 min; final extension at 72 °C for 10 min. PCR products were cloned in *Escherichia coli* TOP10 cells using a TA cloning kit (Invitrogen Life Technologies) following the manufacturer's protocol. Plasmid DNA was purified using the QIAprep Miniprep (Qiagen). Cloned fragments were sequenced commercially on both strands using vector (M13F and R) and internal (16S.19F and 16S.409R) primers (Lark Technologies).

Phylogenetic analyses used both our novel sequences and sequences available from GenBank (<http://www.ncbi.nlm.nih.gov/>). Because of computational power impediments, maximum-parsimony and maximum-likelihood analyses mostly included only complete sequences of *ssu rDNA*; where included, short sequences were located in the most variable region of the gene. Taxa for inclusion in analyses were selected to provide a broad representation of phylogenetic diversity from the freshwater, brackish, marine and Antarctic lineages of the Syn/Pro clade (Sánchez-Baracaldo *et al.*, 2005). To avoid redundancy, we excluded sequences that showed high pair-wise similarities within previously reported clades (Crosbie *et al.*, 2003; Ernst *et al.*, 2003; Fuller *et al.*, 2003). Sequences were aligned manually according to their predicted secondary structure using Se-Al v2.0a11 (Rambaut, 1996); only characters that could be unambiguously aligned were used. Sequence alignments were exported as NEXUS files and analysed using unweighted maximum-parsimony (MP), maximum-likelihood (ML) and neighbour-joining (NJ) in PAUP* (Swofford, 2001). In addition, bootstrap analyses were performed to evaluate branching topologies. For the ML analyses, the estimation of parameters and choice of best-fit model of nucleotide substitution for the data were conducted using the computer program Modeltest 3.0 (Posada & Crandall, 1998), using hierarchical likelihood ratio tests. As a result, the GTR + G + I model of nucleotide substitution was used. These parameters were then used in optimal tree searches, which consisted of 100 random sequence additions and TBR (tree-bisection-reconnection) branch swapping. Clade support was estimated by 10 000 bootstrap replicates with 10 random-addition sequence replicates per bootstrap replicate. Unlike in previous analyses (Ferris & Palenik, 1998; Rocap *et al.*, 2002; Crosbie *et al.*, 2003) *Synechococcus elongatus* PCC 7942 and/or *Synechococcus* PCC 6301 were not used to root the trees because phylogenomic analyses have shown that *S.*

elongatus is probably not the sister group of the Syn/Pro clade (Sánchez-Baracaldo *et al.*, 2005). It has been shown that the ABRAXAS and ACE *Synechococcus* isolates from the Vestfold Hill Lakes, eastern Antarctica (Vincent *et al.*, 2000), are sister to the marine and freshwater lineages within the Syn/Pro clade (Sánchez-Baracaldo *et al.*, 2005); therefore, these strains were used when rooting.

Quantitative PCR. Four distinct freshwater picocyanobacterial clades were selected for quantitative Taq nuclease assays (clades I–IV; Fig. 1). Primers (Table 2) were designed to amplify the most variable region of the *ssu rDNA* (between 721 and 811 bp from the 5'-end of the encoded *ssu rRNA* molecule) using Primer Express software version 2.0 (PE Applied Biosystems); for each primer, the number of mismatches with non-target sequences varied between 5 and 7 bp (Table 2). We used two MGB probes (Table 2), one, P100, for picocyanobacterial clades I, II and III and one, P104, specific for clade IV. Both probes were labelled with a 5' FAM reporter and a non-fluorescent quencher at the 3'-end; the minor groove-binding activity of the quencher at the 3'-end increases the melting temperature (*T_m*), allowing the use of shorter probe sequences (Applied Biosystems). For initial specificity tests and for the construction of standard curves, known concentrations of cloned 390 bp *ssu rDNA* fragments of the following samples were used: cwp31 (clade I); AP10 8 (clade II); cwp9 1 (clade III) and cwp124 3 (clade IV) (Fig. 1). Primer specificity for each clade was tested using SYBR Green to follow product accumulation in quantitative PCR: 10 µl reactions contained 1 µl of 2 × SYBR Green buffer (with ROX as a passive reference for normalization, Applied Biosystems), 0.3 µl of 10 µM primer (MWG-Biotech), 1 µl of 20 mM dNTP, 1.2 µl of 25 mM MgCl₂, 0.005 µl AmpliTaq Gold DNA polymerase, 2.15 µl water and 4 µl template DNA. For these reactions the template comprised either a single cloned *ssu rDNA*, or mixtures of cloned sequences where the target was present at a concentration of about 10⁵ molecules ml⁻¹ in the presence of each of the three non-target sequences either at the same concentration or at 10 × or 0.1 × relative concentrations. All real-time quantitative PCRs used a 96-well plate format and an ABI 7000 sequence detection system (PE Biosystems) as follows: 50 °C for 2 min, 95 °C for 10 min; 35 cycles of 95 °C for 15 s and 60 °C for 1 min. Quantitative TaqMan assays were performed in a final reaction volume of 17.5 µl. Reactions used to construct standard curves contained 1.57 µl of each primer (10 µM,

MWG-Biotech), 0.7 µl probe (5 µM, Applied Biosystems), 8.8 µl 2 × TaqMan Universal PCR Master Mix (containing AmpliTaq Gold DNA polymerase, AmpErase UNG, dNTPs with dUTP, passive reference 1, and optimized buffer components; AB Applied Biosystems), 2.86 µl water and 2 µl of the diluted standard template DNA (a five-step 10-fold dilution series providing approximately 10⁷ to 10³ molecules per reaction: actual concentrations differed for the individual clones). In negative controls additional water replaced the template DNA. The molar concentration of template used for the construction of standard curves was calculated from the DNA concentration, determined from the A₂₆₀, and the molar mass of the recombinant plasmid carrying the target sequence. For the analysis of natural microbial communities, amplification reactions contained primer, probe and buffer (as above) together with 2.8 µl environmental DNA, 2 µl of the most dilute target DNA standard and 0.06 µl water; the dilute DNA standard was added to each reaction to ensure that all measured template concentrations fell within the range of the standard curve. Cycling parameters were as described above. For all samples and standards, four replicate amplification reactions were used. C_t values for all standards and environmental samples were obtained using the ABI Prism 7000 SDS software, version 1.1 (Applied Biosystems) and the data were exported into Excel, where

template concentrations in environmental samples were calculated using the standard curves and from the known proportion of the environmental DNA sample added to the amplification reaction.

RESULTS AND DISCUSSION

Phylogenetic analyses

Phylogenetic analyses of the sequences obtained from freshwater and Baltic Sea picocyanobacteria (Table 1; Fig. 1) suggest a greater diversity of picocyanobacteria in freshwater than marine environments, which is in keeping with the results of Crosbie *et al.* (2003). These results are consistent either with a long evolutionary history for freshwater picocyanobacteria or with more rapid speciation associated with the geographical barriers created by the fragmented nature of some freshwater systems, such as glacier-derived lakes. The relative homogeneity within the marine picocyanobacteria could be explained by the

Fig. 1. Neighbour-joining phylogenetic tree of ssu rDNA sequences from unicellular picocyanobacteria, Syn/Pro clade *sensu* Sánchez-Baracaldo *et al.* (2005). Bootstrap values (1000 replicates) are shown above nodes; values lower than 50 are not shown. Terminal branches indicate strain names and accession numbers. Names and numbers in bold indicate strains for which ssu rDNA was determined in this study: Abbot's Pool (AP), Baltic Sea (BS), Cotswold Water Park (CWP) and Esthwaite Water (EW), with names for short sequences in lower case, and names for long sequences in upper case. Clade composition was confirmed by maximum-parsimony and maximum-likelihood.

connectivity between different ocean basins giving rise to global populations, although some authors argue very strongly that even planktonic marine micro-organisms display distinct biogeographic differentiation (Medlin, 2007). Our results agree with other analyses in showing marine *Synechococcus* as a sister group to *Prochlorococcus*, which together form a monophyletic clade, sharing a relatively recent common ancestor (Honda *et al.*, 1999; Rocap *et al.*, 2002; Litvaitis, 2002). Both the marine *Synechococcus* and *Prochlorococcus* are themselves natural groups (Rocap *et al.*, 2002).

Sequences derived from samples collected from freshwaters fall into a number of both novel and previously described clades (Fig. 1). We recovered similar clades to previous studies of the non-marine picocyanobacteria (Ernst *et al.*, 2003; Crosbie *et al.*, 2003). Cultured isolates from Esthwaite Water and an environmental sequence from Cotswold Water Park (CWP) Lake 31 are part of the *Cyanobium gracile* cluster (clade I, Fig. 1). A novel cluster composed of environmental sequences from Abbot's Pool, Esthwaite Water and CWP Lakes 9, 123 and 124 form a well-supported group (clade II, Fig. 1) where BLAST searches failed to identify any closely related sequences in GenBank. Environmental sequences from Priest Pot and CWP Lake 9 form another well-supported clade (clade III, Fig. 1), again with no similar sequences found in GenBank. Finally, environmental sequences from CWP Lake 124 formed a cluster with sequences from Mondsee (Austria) and Lake Biwa (Japan) (clade IV, Fig. 1) that corresponds to group H described by Crosbie *et al.* (2003). Of the novel groups reported here, clades II and III, clade II not only seems to be more diverse in clone libraries than clade III, it also seems to be more widely distributed, being found in several lakes. Organisms from within the *C. gracile* cluster seem to be abundant and widespread. Our results show that most previously described clusters also contain geographically widespread isolates, and only a few are limited to a single location, a finding similar to that reported by Crosbie *et al.* (2003), who found, for example, cluster B, identified from deep subalpine lakes, also encompassed isolates from several Japanese lakes and Lough Neagh in Ireland.

The three cultured picocyanobacterial isolates from the Baltic Sea do not form a monophyletic group, but cluster with different clades (Fig. 1). B9801 falls within the *C. gracile* cluster (Ernst *et al.*, 2003; Crosbie *et al.*, 2003) that includes a previously described isolate from the Baltic Sea (BS20, Ernst *et al.*, 2003), B9802 clusters with freshwater sequences from Mondsee or group I (Crosbie *et al.*, 2003) that forms a sister group to both marine and other freshwater lineages within the Syn/Pro clade, and B9803 forms part of a cluster that includes the halotolerant WH5701 assigned to *Synechococcus* subcluster 5.2 (Fuller *et al.*, 2003; Fig. 1). Sequences derived from Baltic Sea environmental DNA samples were also distributed among both marine and freshwater clusters. Sequence bsL28-2 (Table 1, Fig. 1) falls within the marine *Synechococcus* cluster, with most other brackish environmental sequences grouping within what was previously

described as subalpine cluster I (Ernst *et al.*, 2003) or cluster B (Crosbie *et al.*, 2003). Interestingly, most Baltic Sea-derived sequences appear more closely related to freshwater lineages, with only one of marine origin. The presence in the Baltic Sea of picocyanobacteria related to freshwater and marine forms is perhaps not surprising given its current brackish state and its changing salinity over the last 12 000 years (Andrén *et al.*, 2000; Bianchi *et al.*, 2000).

Quantitative analysis of picocyanobacterial community structure

For community structure analysis using TaqMan assays we selected four well-supported clades from our phylogenetic analyses (Fig. 1). Oligonucleotide probes and primers were designed to variable regions within the sequenced region of the *ssu* rDNA for each target clade (Table 2). Such variable regions were identified using an *ssu* rDNA sequence alignment that included 58 taxa from the Syn/Pro clade and representatives of the main cyanobacteria clades shown by Sánchez-Baracaldo *et al.* (2005). Because there was only a small variable region available for probe and primer design, it was more efficient to use the shorter MGB probes (Table 2). Probe P100 was designed to target *ssu* rDNAs of clades I–III, with P104 specific for clade IV (Fig. 1); differentiation between clades I–III was achieved through the specificity of the amplification primers (Table 2). Although there is only one mismatch between the two probes, P100 cannot be used to identify clade IV targets and probe P104 cannot be used to identify clades I–III. Probe specificity was confirmed using cloned *ssu* rDNA fragments for each target clade (data not shown).

Previous studies of picocyanobacterial community structure have made use of multiple subsamples extracted from membrane filters (Becker *et al.*, 2000, 2002, 2004). To test whether a single subsample taken from a population sample collected on a filter is representative of the whole, we extracted DNA from three separate punch subsamples from each of eight replicate filters and then determined the concentration of clade IV template in each (Table 3). The results demonstrated that C_t values for individual amplification reactions can deviate from the mean for that sample by a maximum of 1.07 cycles ($=1.9 \times$ under- or overestimation of the initial template concentration), but that overall the measured deviation in C_t is likely to be <0.31 (± 0.05) cycles ($=1.2 \times$ under- or overestimation of the initial template concentration). From these measurements we feel confident that single DNA extracts from population samples are only likely to under- or overestimate the abundance of particular organisms by a maximum of about twofold.

The ability of primer/probe combinations to amplify target sequences in the presence of non-target DNA was initially confirmed by amplification of cloned *ssu* rDNA sequences from mixtures of equimolar, $10 \times$ excess and $10 \times$ less concentrated non-target sequences. In all cases the quantitative PCR-based estimates of template concentration were

Table 3. Quantitative PCR determination of the abundance of clade IV ssu rDNA from replicate samples of a single population

Replicate	Overall mean C_t *	Max. individual deviation from the mean C_t (cycles)†	Mean deviation from the overall mean C_t (cycles)‡
1	24.97	0.72	0.24 (± 0.11)
2	24.58	0.78	0.50 (± 0.16)
3	25.61	0.77	0.42 (± 0.11)
4	25.80	0.23	0.12 (± 0.04)
5	26.48	0.36	0.16 (± 0.06)
6	23.74	1.07	0.68 (± 0.15)
7	25.82	0.25	0.12 (± 0.04)
8	24.39	0.35	0.21 (± 0.06)
Total			0.31 (± 0.05)

*Calculated from the 12 individual readings, i.e. four PCR replicates from each of three DNA extractions.

†Maximum value by which an individual assay deviates from the overall mean.

‡Numbers in parentheses are 95 % confidence intervals.

unaffected by the presence of non-target sequences and recovered the expected value for the target sequence concentration. Furthermore, the C_t values obtained for the internal standards added to all environmental samples did not change even in the presence of up to a 70-fold excess of non-target template (data not shown).

TaqMan assays to quantify community structures

The concentration of the picocyanobacterial ssu rDNA genes in natural communities was estimated over an annual cycle for CWP Lakes 9 and 124 during 2003/4, and the Baltic Sea during July 2003. Community DNA extracted from water samples of known volume provided the template for clade-specific assays where the release of the fluorescent dye from the TaqMan probe was used to quantify the accumulation of amplification products. Comparisons of the rate of product accumulation, and hence initial template concentration within community DNA samples, were achieved by evaluating the number of amplification cycles (C_t) needed for the released fluorescence to reach a threshold value. The efficiency of amplification for the environmentally derived samples and the calibration standards did not differ significantly within a single PCR run, but there was significant variation between runs (data not shown). The inclusion of standards in all runs was essential for robust estimates of template concentrations in the samples derived from nature. Absolute quantification of initial template concentrations was achieved using calibration curves generated from the C_t values measured for cloned clade-specific templates of known ssu rDNA concentration. All environmental samples were spiked with the most dilute standard, equivalent to a template concentration of between 854 and 1651 ssu rDNA fragments per reaction, to ensure that amplification reactions gave C_t values falling within the range of the calibration curves. The quantification and detection

limit of these assays was dependent on obtaining a C_t value significantly lower than that for the most dilute standard, which in practice was found to equate to a template concentration of about 100–200 molecules per reaction. From the measured concentration of ssu rDNA molecules in individual assays, the concentration in the original water samples was calculated from (a) the proportion of the total extracted DNA used in the assay (2.8 μ l from 400 μ l; correction factor is $\times 142.9$), (b) the area of the filter from which the DNA was extracted as a proportion of the total filter area on which the community sample was collected (50.3 mm² from 1134.1 mm²; correction factor is $\times 22.6$), and (c) the volume of water filtered. Using this method we were able to follow the development of populations of individual picocyanobacterial clades over an annual cycle (Fig. 2) and to quantify community structures at individual sampling locations in the Baltic Sea (Table 4).

The concentration of picocyanobacteria from all clades in lake surface waters varied over an annual cycle (Fig. 2). Organisms from clades II, III and IV could not be detected in samples collected over winter. For clade I in CWP Lake 9 there were three distinct periods of population development, one in the spring (April/May), one in the summer (July), and a much smaller population maximum in the autumn (September/October). Organisms from clades II, III and IV developed marked population maxima in the summer (July/August) but with some indication of a period of enhanced growth in the autumn (September/October). Clade I could be detected in the water of CWP Lake 124 throughout the year, never falling below a concentration of 3000 ssu rDNA molecules ml⁻¹. The periods of rapid increase suggest minimum population doubling times of between 0.86 and 1.26 days for clade I, 0.91 days for clade II, 1.94 days for clade III and 4.92 days for clade IV; the minimum halving times for the declining populations were 0.86–1.03 days for

Fig. 2. Abundance of clades I–IV for 2003/4 obtained by absolute quantification using TaqMan assays based on ssu rDNA sequences.

clade I, 0.92 days for clade II, 1.18–1.41 days for clade III and 3.71 days for clade IV. Although the trends in population abundance observed here are likely to be robust, the exact numerical values need to be interpreted with

caution. Previous studies have demonstrated losses of picocyanobacteria when passing natural samples through a range of pre-filters with different pore sizes (Becker *et al.*, 2002), and thus it is possible that abundances shown (Fig. 2)

Table 4. Abundance of picocyanobacteria from clades I to IV in samples of communities collected from the Baltic Sea in July 2003

Station position	Date*	Clade I ssu mol ml ⁻¹	Clade II ssu mol ml ⁻¹	Clade III ssu mol ml ⁻¹	Clade IV ssu mol ml ⁻¹
59°35'N 23°37'E	14/07/2003	4.0 × 10 ⁴	3.7 × 10 ³	5.2 × 10 ³	4.9 × 10 ⁴
59°41'N 23°31'E	15/07/2003	3.2 × 10 ⁴	ND	ND	3.0 × 10 ⁴
60°04'N 26°21'E	16/07/2003	9.7 × 10 ⁴	2.7 × 10 ³	ND	3.3 × 10 ⁴
59°51'N 24°50'E	17/07/2003	8.4 × 10 ⁴	3.5 × 10 ³	ND	2.3 × 10 ⁴
59°35'N 23°18'E	18/07/2003	3.7 × 10 ⁴	ND	ND	1.3 × 10 ⁴

ND, Below the limit of detection.

*Day/month/year.

might underestimate natural population sizes. Furthermore, concentrations of ssu rDNA are not equivalent to cell concentrations, because the ssu rDNA copy number per cell is known to be variable (Herdman *et al.*, 1979; Labarre *et al.*, 1989). No attempt was made during the current study to quantify picocyanobacterial abundance by direct cell counts or flow cytometry.

Picocyanobacteria from temperate lakes of all trophic types have seasonal cycles and a variety of successional patterns (Weisse, 1993). While some studies of temperate lakes have shown two peaks of picocyanobacteria abundance, one in spring and one in late summer (Weisse, 1993; Callieri & Stockner, 2000), other lakes lack a spring peak, with only a summer or autumn population maximum (Pick & Agbeti, 1991; Maeda *et al.*, 1992; Hawley & Whitton, 1991). Our data suggest that CWP lakes 9 and 124 exhibit one peak, in summer, for all clades with the exception of clade I. It is difficult to interpret the significance of these observations as we know nothing about the characteristics of the organisms studied here, although differences in apparent doubling times suggest that they may encompass a range of physiologies. Remarkable physiological differences have been demonstrated between closely related Baltic Sea *Synechococcus* strains that only exhibit 1% sequence divergence in their rDNA ITS-1 (Stomp *et al.*, 2004); therefore, the phylogenetic groupings based on the more highly conserved ssu rDNA sequences may be indicative of a great deal of physiological diversity. Moreover, our results show that there are temporal differences in population development among clades, with considerable differences in the maximum observed concentration of ssu rDNA molecules (Fig. 2, Table 4).

Interestingly, there were differences when comparing cloning and quantitative PCR experiments. PCR-based analyses over an annual cycle detected ssu rDNA molecules from clades that were not identified in clone libraries prepared from the same water body. This is perhaps not surprising and provides an illustration of the need to adopt a number of approaches when attempting to describe diversity in microbial communities. It is also worthy of note that some clades have no cultured representatives, which is almost certainly a function of both the highly variable abundance over an annual cycle and the incubation conditions used for culture establishment.

All freshwater clades studied here seem to form significant populations in the brackish waters of the Baltic Sea. Members of each clade were found at relatively high abundance during the summer of 2003, with ssu rDNA concentrations in the order of 10^3 – 10^4 ml⁻¹ (Table 4). Three of the samples (collected on 14, 15 and 18 July 2003) were from locations <20 km apart, with the remaining sampling locations either about 80 or 170 km from the others. For clades I and IV the observed variation in abundance was probably not significant over these spatial scales, i.e. it is within the twofold error range that could be associated with using just a single DNA extract per sample (see above), but for clades II and III the variation in abundance is significant,

with organisms belonging to these clades being undetectable in some samples, but reaching concentrations between 2.7 and 5.2×10^3 ml⁻¹ in others (Table 4). These results illustrate that even in open water environments the picocyanobacterial community structure is patchy over quite small spatial/temporal scales. This patchiness needs to be borne in mind when interpreting field data related to these fast-growing and rapidly disappearing organisms.

Conclusions

Freshwater picocyanobacteria communities encompass a large number of genetically and evolutionarily distinct lineages. The use of quantitative PCR allows the assessment of the contribution of individual clades (groups) to community structure and shows that their abundance varies significantly over relatively short spatial and temporal scales. The dynamic patterns of abundance and distribution shown here seem to support evidence from previous ecological studies. Molecular ecology studies based on novel clades, such as II and III, provide additional and valuable information on the diversity of picocyanobacteria in temperate lakes. Further studies should focus on factors driving changes in picocyanobacterial community structure and on determining both the physiological diversity and the contribution to ecosystem services associated with each of the different clades.

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